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Switchgrass for Bioethanol and Other Value-Added Applications: A Review

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Abstract

Switchgrass is a promising feedstock for value-added applications due to its high productivity, potentially low requirements for agricultural inputs and positive environmental impacts. The objective of this paper is to review published research on the conversion of switchgrass into bioethanol and other value-added products. Environmental benefits associated with switchgrass include the potential for carbon sequestration, nutrient recovery from runoff, soil remediation and provision of habitats for grassland birds. Pretreatment of switchgrass is required to improve the yields of fermentable sugars. Based on the type of pretreatment, glucose yields range from 70% to 90%, and xylose yields range from 70% to 100% after hydrolysis. Following pretreatment and hydrolysis, ethanol yields range from 72% to 92% of the theoretical maximum. Other value-added uses of switchgrass include gasification, bio-oil production, newsprint production, and fiber reinforcement in thermoplastic composites. Future prospects for research include increased biomass yields, optimization of feedstock composition for bioenergy applications, and efficient pentose fermentation to improve ethanol yields.

Keywords: switchgrass, bioenergy, ethanol, lignocellulosics, pretreatment

1. Introduction

Switchgrass (Panicum virgatum L.) is a perennial warm-season C$_4$ grass native to North America. It occurs naturally from 55°N latitude to central Mexico (Lewandowski et al., 2003). Tolerance to heat, cold, and drought have enabled adapted ecotypes of switchgrass to inhabit regions throughout North America (Casler et al., 2007). Historically, switchgrass was one of the dominant grasses in the North American tall-grass prairie and was adapted to other regions by European settlers (Hitchcock, 1971). Perennial grasses like switchgrass have been widely used for forage purposes in their native state prior to being established as a crop. Since the 1940s, switchgrass has been used for conservation and warm-season pasture purposes in the Great Plains and Midwest states (Vogel, 2004). There are two general ecotypes of switchgrass: lowland ecotypes that are vigorous, tall, thick-stemmed, and adapted to wetter conditions, and upland ecotypes that are shorter, thinner-stemmed, and adapted to drier conditions (Gunter et al., 1996). Switchgrass was identified by the U.S. Department of Energy as a model herbaceous energy crop (McLaughlin, 1996). Switchgrass shows promise due to its high productivity, suitability for marginal land quality, low water and nutritional requirements, environmental benefits, and flexibility for multipurpose uses (McLaughlin et al., 1999). Switchgrass can be easily integrated into existing farming operations because conventional equipment for seeding, crop management, and harvesting can be used (Lewandowski et al., 2003; Vogel et al., 2002).

Research on the feedstock potential of switchgrass has been effectively summarized in previous studies such as those presented by Sanderson et al. (2006) and Parrish and Fike (2005). Traditionally switchgrass has been bred primarily to improve its nutritional value for use as a forage crop. Consequently, breeding strategies typically emphasized high leaf to stem ratio and high nutrient content. However, later studies emphasized the importance of high cellulose content for bioethanol production and low ash content for combustion systems (McLaughlin et al., 1999).

With newer varieties of switchgrass, yields in excess of 20 Mg ha$^{-1}$ have been reported for test plots. For example, Sanderson et al. (1996) reported 26 Mg ha$^{-1}$ in field trials in Texas, and Thompson et al. (2004) reported 36.7 Mg ha$^{-1}$ from field work in Oklahoma. Although these high yields are site-specific for test plots and do not reflect realistic expectations, Parrish and Fike (2005) state that with well-adapted cultivars from breeding and biotechnology research, annual yields of over 15 Mg ha$^{-1}$ are feasible for lands that receive annual rainfall of at least 70 cm with nitrogen applications of 50 kg N ha$^{-1}$ year$^{-1}$. Schmer et al. (2008) reported on-farm yields ranging from 5.2 to 11.1 Mg ha$^{-1}$ for field trials in the mid-continental U.S., which are more realistic expectations for long-term yields. Dunn et al. (1993) identified the benefits associated with the perennial nature of switchgrass—such as less intensive agricultural management practices, reduced energy and agrochemical consumption and positive effects on soil and wildlife quality.
While the conversion of lignocellulosic materials has been previously summarized, a review of the significant amount of research specific to the conversion of switchgrass has not been presented to date. The objective of this paper is to review published research on the conversion of switchgrass into bioethanol and other value-added products in addition to discussing research on potential environmental benefits associated with switchgrass.

2. Environmental benefits

When the U.S. Department of Energy chose switchgrass to be a major focus of research for bioenergy production, among the reasons cited was the expected positive environmental impact as evidenced by results from the conservation reserve program (CRP). The U.S. Congress established this program in 1985 to remediate the negative effects of decades of row-crop production. The program designated 90% of impacted land areas to perennial grasses like switchgrass. Since soil tillage is only required in the establishment year, there is a reduced risk of soil erosion (Ma et al., 2000a). Hohenstein and Wright (1994) estimated a 95% reduction in soil erosion rates and a 90% reduction in pesticide use for herbaceous energy crops such as switchgrass relative to annual row crops like corn and soybean. While, the initial establishment of harvestable stands of switchgrass requires the use of high quality seed and herbicides for weed control, subsequent stand maintenance requires only limited and periodic herbicide applications (Sarath et al., 2008).

Bransby et al. (1998) stated that switchgrass could reduce CO₂ emissions and improve soil quality by carbon sequestration. In addition to the large amount of above-ground biomass, switchgrass has an extensive and deep root system that is beneficial for increasing soil carbon storage in addition to being at least 50% more effective in water use compared to cool-season grasses (Stout et al., 1998). The root system can account for up to 80% of the total biomass (Liebig et al., 2005). Garten and Wullschleger (2000) presented a model that predicted a 12% increase in soil organic carbon inventory over a 10-year period following switchgrass establishment with experimental data indicating that the majority of increase is observed in the first 10 cm of the soil. Gebhart et al. (1994) reported that perennial grasses like switchgrass can store 1.1 Mg of carbon per hectare annually in the upper 1 m of the soil on CRP lands. Bransby et al. (1998) stated that these studies on CRP lands are probably not representative of switchgrass grown for energy production since the emphasis would shift from conservation to maximizing yields. However, McLaughlin et al. (2002) indicated that switchgrass grown on bioenergy research plots could add 1.7 Mg of carbon per hectare annually, which is higher than CRP estimates reported by Gebhart et al. (1994). Sanderson (2008) reported a 33% increase in soil carbon in the 0–5-cm layer and a slight decrease in the 5–15 cm and 15–30 cm layers seven years after planting. A systematic study of carbon dynamics following the establishment of switchgrass showed that over a two-year period, the top 15 cm of sandy loam soil exhibited a 12% increase in carbon mineralization, a 168% increase in microbial biomass carbon, and a 116% increase in net carbon turnover (Ma et al., 2000b).

Wu et al. (2006) estimated that the net amount of CO₂ sequestered would be around 48.5 kg per dry metric ton of switchgrass. Frank et al. (2004) examined Sunburst and Dacotah switchgrass cultivars and noted that the net system carbon gain doubled over a three-year period. Combined with the zero net carbon exchange as a result of burning bioethanol from switchgrass, addition of soil carbon results in the overall reduction of atmospheric release of CO₂ (Lynd et al., 1991). However, such gains as a result of carbon sequestration are not guaranteed. Bransby et al. (1998) noted that switchgrass will provide net gains in C sequestration when grown on soil with a history of annual row crops and not when it replaces grazed pasture. McLaughlin et al. (2002) also predicted smaller gains in soil carbon sequestration following conversion of pastures to switchgrass and noted that significant gains can be achieved for highly degraded soils in warm climates.

The use of switchgrass to improve surface water quality has also been examined. Lee et al. (1998) compared switchgrass filter strips to cool-season grass filter strips and reported that switchgrass was more effective in removing phosphorous and nitrogen from runoff. Sanderson et al. (2001) treated a switchgrass filter strip with dairy manure and noted similar results for phosphorous and nitrogen reduction. They also noted a 40–44% reduction in chemical oxygen demand as a result of the filter strip. Mersie et al. (1999) utilized switchgrass filter strips to reduce the amount of dissolved herbicides (atrazine and metachlor) by 52% and 59%, respectively. These reductions were attributed to slower runoff velocities and increase in soil retention. Mersie et al. (2006) reported significant levels of sorption of atrazine and metachlor to switchgrass residues that accumulate over time on filter strips.

Entry and Watrud (1998) tested the ability of Alamo switchgrass to remediate soil contaminated by cesium-137 and strontium-90. These elements are radionuclides released during nuclear testing, nuclear reactor accidents, and weapons production. The authors reported a 36% removal of cesium and a 44% removal of strontium over a five month period.

Switchgrass also has positive impacts on wildlife by providing a suitable habitat for grassland birds that are rapidly declining in numbers (Murray et al., 2003). Switchgrass is typically harvested in late summer or the fall. By this time, the breeding season for most grassland birds is over and there is minimal disturbance to nesting birds (Roth et al., 2005). Murray and Best (2003) proposed a balance between harvested and non-harvested switchgrass fields to preserve species richness of grassland birds. While most studies suggest that switchgrass on CRP lands offers the best habitat for grassland bird species, the scenario is short-term, as these lands are typically in the program for a 10-year period. Switchgrass grown on lands for biofuel harvesting could offer a longer-term habitat given that they will be tied into the life of an energy production facility (Roth et al., 2005).

3. Bioethanol production

With the steady increase in energy consumption and current dependence on crude oil to meet energy demands, there is considerable and immediate interest in developing alternative energy sources. Campbell and Laherrere (1998) predicted a steady decline in crude oil production in the coming decades and stated that the world is already in a peak-oil situation. More optimistic estimates predict a peak-oil situation within the next two decades (Wood et al., 2003). Bioethanol is a promising alternative to reduce dependence on crude oil. Currently, the U.S. produces most of its bioethanol from corn grain. This may not be a viable long-term option as increasing demand for corn-based ethanol will have significant arable land requirements (Sun and Cheng, 2002). Increased use of corn for ethanol production will result in higher corn prices and will negatively impact the food and feed industries and could result in reduced exports of animal products (Elobeid et al., 2007). In lieu of these issues, lignocellulosic feedstocks have been considered for ethanol production. These materials include agricultural residues, cellulosic waste such as newsprint and office paper, and herbaceous and woody crops. In particular, herbaceous energy crops such as switchgrass are viewed as a potential long-term ethanol feedstock to replace corn.

Morrow et al. (2006) state that a mature bioenergy crop production system would yield 330–380 I of ethanol per Mg of dry switchgrass. These estimates are consistent with those from the National Renewable Energy Laboratory’s theoretical ethanol yield calcula-
tor (http://www.nrel.gov/biomass/energy_analysis.html) that assumes conversion of both hexoses and pentoses. Using a biomass yield of 15 Mg ha⁻¹ (Parrish and Fike, 2005), the corresponding theoretical ethanol production potential from highly adapted switchgrass varieties is between 5000 and 6000 l ha⁻¹. With more realistic biomass yields reported by Schmer et al. (2008), theoretical ethanol yields are likely to be between 2000 and 4000 l ha⁻¹. In comparison, theoretical ethanol yields from corn starch is approximately 4000 l ha⁻¹, based on data from Gulati et al. (1996) and 2000 l ha⁻¹ from corn stover, based on data from Perllack and Turhollow (2003). However, Perllack and Turhollow noted that the actual yields from corn stover would be much lower (~700–800 l ha⁻¹) since a large portion of the stovers must be left in the fields for maintenance of soil quality.

The major components of lignocelluloses are cellulose, hemicellulose, and lignin, which are closely associated in a complex crystalline structure. Table 1 shows the amount of cellulose, hemicellulose, and lignin in different switchgrass varieties in the National Renewable Energy Laboratory’s biomass feedstock properties and composition database available online (http://www.nrel.gov/biomass/energy_analysis.html). Figure 1 describes the general process for converting the carbohydrates in switchgrass into ethanol. Unlike corn, where starch carbohydrates are easily depolymerized into fermentable sugars, carbohydrate fractions in lignocelluloses (cellulose and hemicellulose) are not readily available for hydrolysis. The efficiency of enzymatic hydrolysis is reduced due to limited accessibility of the enzymes to cellulose (McMillan, 1994). Hence, as shown in Figure 1, pretreatment is required to improve accessibility of enzymes to cellulose and hemicellulose fractions. Following pretreatment, cellulose and hemicellulose fractions can be hydrolyzed into fermentable sugars, while lignin can be recovered and used as a fuel to meet some of the energy requirements in a bioethanol production system (Wyman, 1994). After hydrolysis, the resulting sugars are fermented into ethanol, which is then distilled for fuel purposes. Currently, there are technological and economic limitations to ethanol production from lignocelluloses in each step in the conversion process shown in Figure 1. Cost effective pretreatment processes, cheaper hydrolytic enzymes, and fermentation of pentose sugars have been identified in this regard (Gray et al., 2006; Wyman, 2003).

### 3.1. Pretreatment processes

The purpose of pretreatment processes is to reduce crystallinity of cellulose, increase porosity of the biomass, and achieve the desired fractionation (Sun and Cheng, 2002). Effective pretreatments must improve enzymatic hydrolysis, minimize carbohydrate losses, and prevent formation of by-products that might inhibit subsequent hydrolysis and fermentation steps. Physical, physico-chemical, chemical, and biological processes have been studied extensively for the pretreatment of lignocellulosic materials, and detailed descriptions of these processes have been previously described by Mosier et al. (2005), Sun and Cheng (2002), and Weil et al. (1994). This section briefly describes the major types of pretreatment and reviews research related to switchgrass in each case.

#### 3.1.1. Physical pretreatment

Physical pretreatment of lignocelluloses typically involves size comminution by grinding, milling, or chipping. The goal is to reduce the crystallinity of the cellulose fibers in the biomass. Size reduction of lignocelluloses is also necessary to eliminate mass and heat transfer limitations during the hydrolysis reactions (Schell and Harwood, 1994). The size of the resulting materials is typically 10–30 mm after chipping and 0.2–2 mm after milling or grinding (Sun and Cheng, 2002).

Bridgeman et al. (2007) studied the effects of size reduction of switchgrass achieved via ball milling. The study reported that for particle sizes smaller than 90 μm, cellulose content was 13.4% lower than for larger sized particles. The losses in lignin and hemicellulose were appreciably less (3.43% and 4.74%, respectively). These results indicate that extensive size reduction is undesirable as it causes significant carbohydrate losses which ultimately results in less reducing sugars and reduced ethanol yield.

Mani et al. (2004) examined the energy requirements for size reduction of switchgrass using a hammer mill. At 8% moisture content, energy requirements increased linearly as particle size reduced and at 12% moisture content, the energy requirements tended to level off for particle sizes less than 2 mm. The study noted that energy requirements for size reduction of switchgrass were higher than that of corn stover, barley straw, and wheat straw at the same moisture content. Additionally, it was reported that higher moisture content biomass required greater energy inputs for size reduction. Yu et al. (2006) reported that size reduction of switchgrass based on shear stress is more efficient than size reduction based on tensile stress and noted that shear stress of switchgrass was less affected by moisture content than tensile stress. Igathinathane et al. (2008) reported a significant increase in cutting energy requirements for switchgrass with high moisture (51%) in comparison to low moisture (9%) using a shear based knife grid for size reduction using a linear grid of knives. Table 2 presents estimates of energy consumption from studies on the size reduction of switchgrass.

#### 3.1.2. Physico-chemical pretreatment

The three types of physico-chemical pretreatments discussed in literature are steam explosion, ammonia fiber explosion (AFEX),...
and CO₂ explosion. In steam explosion, size reduced biomass is subjected to high-pressure saturated steam for a short time before a sudden drop in pressure causes an explosive decompression of the biomass (McMillan, 1994). Typical conditions are 160–260 °C and 0.69–4.83 MPa (Sun and Cheng, 2002). The process causes transformation of lignin and degradation of hemicelluloses, which improves the enzymatic hydrolysis of cellulose. Steam explosion is known to be a cost effective pretreatment for hardwoods and agricultural residues. However, the process produces inhibitory by-products that may impact downstream processes (Mackie et al., 1986). AFEX and CO₂ explosion are similar to steam explosion. The biomass is exposed to liquid ammonia or CO₂ at high temperature and pressure for a short period of time, followed by a sudden drop in pressure. Unlike steam explosion, AFEX does not solubilize hemicelluloses (Vlasenko et al., 1997) but does require recovery of the ammonia for cost and environmental reasons (Sun and Cheng, 2002). CO₂ explosion is not as effective as AFEX or steam explosion (Dale and Moreira, 1982).

To date, AFEX is the only physico-chemical pretreatment that has been applied for the pretreatment of switchgrass to improve enzymatic hydrolysis. Dale et al. (1996) reported a 5–8-fold improvement in reducing sugar yields over untreated samples after a 48-h enzymatic hydrolysis of AFEX pretreated switchgrass. These results were similar to a more comprehensive study presented by Alizadeh et al. (2005) that optimized AFEX pretreatment of switchgrass by examining the impact of ammonia loading, moisture content, and reactor temperature on the efficiency of enzymatic hydrolysis. The authors reported optimum pretreatment conditions of 100 °C reactor temperature, ammonia loading of 1 g g⁻¹ of biomass and a residence time of 5 min. These conditions yielded a 6-fold improvement in hydrolysis efficiency.

### 3.1.3. Chemical pretreatment

Chemical pretreatment of lignocelluloses includes the use of ozone, acids, alkali, organic solvents, and peroxides. Ozonolysis is carried out at room temperature and is effective at lignin removal without the formation of toxic by-products (Vidal and Molinier, 1988). However, the large ozone requirement makes the process expensive (Sun and Cheng, 2002). Dilute acid pretreatment with sulfuric acid has been studied extensively and is efficient at removal of hemicelluloses but fails to effectively remove lignin. Unlike steam explosion, AFEX does not solubilize hemicelluloses (Vlasenko et al., 1997) but does not require recovery of the ammonia for cost and environmental reasons (Sun and Cheng, 2002). CO₂ explosion is not as effective as AFEX or steam explosion (Dale and Moreira, 1982).

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### Table 2. Energy requirements of size reduction of switchgrass from published studies

<table>
<thead>
<tr>
<th>Reference</th>
<th>Energy consumption (kWh/t)</th>
<th>Moisture content (%)</th>
<th>Size reduction</th>
</tr>
</thead>
<tbody>
<tr>
<td>Samson et al. (2000)</td>
<td>44.9</td>
<td>Not reported</td>
<td>5.6 mm pellets</td>
</tr>
<tr>
<td>Jarusach et al. (2001)</td>
<td>55.9</td>
<td>10–12</td>
<td>2.8 mm screen size</td>
</tr>
<tr>
<td>Mani et al. (2004)</td>
<td>27.6</td>
<td>8</td>
<td>3.2 mm screen size</td>
</tr>
<tr>
<td>Mani et al. (2004)</td>
<td>23.8</td>
<td>12</td>
<td>3.2 mm screen size</td>
</tr>
<tr>
<td>Ightinathane et al. (2008)</td>
<td>2.4–3.2</td>
<td>9</td>
<td>25.4 mm knife grid spacing</td>
</tr>
<tr>
<td>Ightinathane et al. (2008)</td>
<td>2.7–5.4</td>
<td>54</td>
<td>25.4 mm knife grid spacing</td>
</tr>
</tbody>
</table>

examined the use of ammonia water and hydrated lime. Ladisch et al. (1978) showed that pretreatment of lignocelluloses with ca- dioxen, a cellulose solvent, resulted in 90% conversion of cellulose into glucose. The use of organic solvents such as methanol, ethanol, acetone, and ethylene glycol, along with inorganic and organic acids as catalysts, has also been studied (Wood and Saddler, 1988; Chum et al., 1988; Thring et al., 1990).

Dilute acid pretreatment of switchgrass for bioethanol production was first examined by Wyman et al. (1992). The pretreatment was conducted at 140 °C for 1 h using sulfuric acid at low concentrations (up to 0.5% v/v). Enzymatic hydrolysis of the resulting biomass yielded up to 70% conversion of cellulose into glucose over a five-day period. Since the removal of hemicellulose during dilute acid pretreatment correlates with enhanced reactivity of cel- lulose to enzymes (Torget et al., 1990), several studies have used percentage of xylose recovery as a means of optimizing dilute acid pretreatments. Wu and Lee (1997) used a two-stage dilute sulfuric acid pretreatment with an acid concentration of 0.0785% (w/w) to successfully remove 100% of hemicellulose from switchgrass. Es- teghlalian et al. (1997) determined that up to 90% of xylose was re- covered as a result of pretreatment at 180 °C with an acid concentra- tion of 1.2% (w/w). It was also noted that temperature was a more significant parameter than acid concentration. Fenske et al. (1998) reported a 96% xylose recovery in a similar study. Chung et al. (2005) reported optimal conditions of 1.2% (w/w) at 180 °C for sulfuric acid pretreatment of switchgrass and determined that 90% of cellulose in the pretreated biomass was converted into re- ducing sugars during a 72 h enzymatic hydrolysis. Dien et al. (2006) reported yields greater than 80% for both glucan and non- glucan sugars for dilute sulfuric pretreatment of switchgrass with an optimal acid concentration of 1.2% (w/v). Although pretreatment temperature was not optimized, the study indicates higher conversion rates at 150 °C in comparison to 120 °C.

Dilute acid pretreatment of switchgrass does not significantly impact lignin removal. Wu and Lee (1997) noted that high lignin content could lead to increased enzyme consumption due to irreversible adsorption of cellulase enzymes to lignin. Hence, they proposed combining dilute acid pretreatment with an ammonia percolation step to remove lignin. Their results indicate that with this additional step, 20% more lignin can be removed to signifi- cantly increase enzymatic digestibility of the pretreated biomass. Kong et al. (1992) noted that alkalis remove acetate groups from hemicellulose, which result in reduced steric hindrance to enzyme molecules in addition to saponification of ester groups that cross- link lignin and hemicelluloses.

Kim and Lee (1996) reported up to 99% delignification using a combined ammonia–hydrogen peroxide percolation pretreatment at 170 °C. They reported reagent loadings of 0.28 g loading of hy- drogen peroxide g⁻¹ biomass, and 10% (w/w) of ammonia. Their results indicated that the enzymatic digestibility of the resulting pretreated biomass was higher than that of pure alpha cellulose. This could be a sign of significant reduction in crystallinity and in- crease in porosity of cellulose in the pretreated biomass. Kurakake et al. (2001) used ammonia water (25–28%) for the pretreatment of switchgrass (2 ml/g biomass) for 20 min at 120 °C and noted a 3–5-fold improvement in reducing sugar production over untreated switchgrass after a 24-h enzymatic hydrolysis. Isci et al. (2008) pre- treated switchgrass in 30% aqueous ammonium hydroxide with liquid–solid ratios of 5 and 10 ml/g and residence times of 5 and 10 days. The study reported a 40–50% reduction in lignin content and 50% reduction in hemicelluloses content as a result of pretreatment.

Lime (calcium hydroxide) pretreatment of switchgrass was in- vestigated by Chang et al. (1997). With a lime loading of 0.1 g g⁻³ of dry switchgrass, a pretreatment time of 2 h at 100 °C, the study showed that a 72-h enzymatic hydrolysis of pretreated biomass yielded five times higher reducing sugars than untreated switch-
3.2. Enzymatic hydrolysis

As shown in Figure 1, following the pretreatment of lignocelluloses, enzymatic hydrolysis is carried out to break down cellulose and hemicelluloses into fermentable sugars such as glucose and xylose. Strong acids such as sulfuric acid and halogen acids are capable of hydrolyzing a wide variety of lignocelluloses into simple fermentable sugars (Wyman, 1994). However, high acid concentrations and extreme conditions make this approach environmentally and economically unsound (Wright and Dagincourt, 1984). Enzymatic hydrolysis is an environmentally friendly alternative that involves using carbohydrate degrading enzymes (cellulases and hemicellulases) to hydrolyze lignocelluloses into fermentable sugars.

3.2.1. Enzymatic hydrolysis of cellulose

Enzymatic hydrolysis of cellulose is typically carried out by cellulases. Unlike conventional hydrolysis using concentrated acid or alkaline reagents, enzymatic hydrolysis requires mild conditions (pH of 4.5 and temperature of approximately 50 °C). Although cellulases are also produced by several bacterial species such as Clostridium, Cellulomonas, and Bacillus (Bisaria, 1998), fungal cellulases have the best potential for commercial scale use (Duff and Murray, 1996). Cellulases are a complex system of three enzymes that act synergistically to hydrolyze cellulose. The three enzyme components are: 1,4-β-d-glucan glucanohydrolase (EC 3.2.1.3), 1,4-β-d-glucan cellobiohydrolase (EC 3.2.1.91), and β-glucosidase (EC 3.2.1.21) (Ladisch et al., 1983; Wright et al., 1988). These enzymes are commonly referred to as endoglucanase, exoglucanase, and cellobiase, respectively.

Endoglucanase randomly cleaves cellulose chains to form glucose, cellobiose, and cellotriose. Exoglucanase attacks the non-reducing end of cellulose to release cellobiose units. Cellobiase cleaves cellobiose units into fermentable glucose units. Most fungal cellulases are deficient in β-glucosidase activity, which must be supplemented since cellulose accumulation results in cellulase inhibition (Ryu and Mandels, 1980). A cellulase dosage of 10 FPU (filter paper units) per gram of biomass is often used in studies as it enables high glucose yields in 48–72 h (Gregg and Saddler, 1996). However, a range of dosage and hydrolysis conditions have been reported depending on composition of the substrates and pretreatment used. Table 3 summarizes information on cellulose activities and hydrolysis conditions from studies specific to switchgrass.

3.2.2. Enzymatic hydrolysis of hemicelluloses

Complete hydrolysis of xylan involves three main enzymes: endo-β-1,4-xylanase which primarily targets the internal β-1,4 bonds between xylose units, exoylanxanase that releases xylose units and β-xylanolase that releases xylose from xylobiose and short chain xylooligosacharides (Saha and Bothast, 1999). While these enzymes are primarily involved in depolymerization, there are also several ancillary enzymes that are responsible for cleaving side-groups. These include α-L-arabinofuranosidase, α-glucuronidase, acetylxylan esterase, furfural acid esterase, and p-coumaric acid esterase (Saha and Bothast, 1999).

Penicillium capsulatum and Talaromyces emersonii have been identified as microorganisms that have complete enzyme systems that degrade xylan (Filho et al., 1991). Other microorganisms that have been reported as sources for hemicellulose-degrading enzymes are Aureobasidium pullulans (Christov et al., 1997) and several Fusarium species (Saha, 2001, 2002). As in cellulase systems, xylan-degrading systems also exhibit synergism (Bachmann and McCarthy, 1991). While the number of enzymes required for xylan hydrolysis is much greater than for cellulose hydrolysis, accessibility to the substrate is easier since xylan does not form tight crystalline structures (Gilbert and Hazlewood, 1993). To date, no comprehensive effort has been reported to optimize hydrolysis of switchgrass using hemicellulose-degrading enzymes.

3.3. Fermentation

The supernatant from enzymatic hydrolysis of lignocelluloses can contain both hexoses and pentoses (if both cellulose and hemicellulose are hydrolyzed). Depending on the lignocellulose source, the hydrosylate typically consists of glucose, xylose, arabinose, galactose, mannose, fucose, and rhamnose (Saha, 2003). Glucose and xylose are the dominant sugars in the mixture. Saccharomyces cerevisiae and Zymomonas mobilis are capable of efficiently fermenting glucose into ethanol, but are unable to ferment xylose. Other yeasts such as Pachysolen tannophilus, Pichia stipitis, and Candida shehatae are known to ferment xylose into ethanol (Wang et al., 1980; Schneider et al., 1981). Duprez (1994) and Hahn-Hagerdal et al. (1994) noted the difficulties associated with commercial use of xylose fermenting yeasts. These include low ethanol tolerance, difficulty in optimization of fermentation parameters and slow rate of fermentation. An alternative approach is to convert xylose into an isomer called xylulose using xylose isomerase (Chiang et al., 1981; Gong et al., 1981; Jeffries, 1981). Xylulose can then be fermented by traditional yeasts. However, Saha (2003) stated that this approach is not cost effective and indicates that focus should be on development of genetically engineered microorganisms capable of fermenting hexoses and pentoses into ethanol. S. cerevisiae is of particular interest in this regard, and recent reviews detail the efforts to improve pentose fermentation using this microorganism (Chu and Lee, 2007; Hahn-Hagerdal et al., 2007).

In addition to separate hydrolysis and fermentation (SHF), other approaches include direct microbial conversion (DMC) and simul-
Table 3. Cellulase activities and hydrolysis conditions from various studies on switchgrass

<table>
<thead>
<tr>
<th>Reference</th>
<th>Pretreatment</th>
<th>Enzyme activity(^a)</th>
<th>Conditions</th>
<th>Result/yield</th>
</tr>
</thead>
<tbody>
<tr>
<td>Alizadeh et al. (2005)</td>
<td>AFEX</td>
<td>cellulase: 15 FPU g(^{-1}) glucan, celllobiose: 40 IU g(^{-1}) glucan</td>
<td>50 °C, 75 RPM, 168 h</td>
<td>93% glucan conversion, 70% xylan conversion</td>
</tr>
<tr>
<td>Wyman et al. (1992)</td>
<td>dilute acid</td>
<td>cellulase: 26 IU g(^{-1}) biomass, celllobiose: 208 IU g(^{-1}) biomass</td>
<td>45 °C, 150 RPM, 192 h</td>
<td>70% cellulose conversion</td>
</tr>
<tr>
<td>Chung et al. (2005)</td>
<td>dilute acid</td>
<td>cellulase: 60 FPU g(^{-1}) cellulose</td>
<td>50 °C, 68 RPM, 192 h</td>
<td>91.4% cellulose conversion</td>
</tr>
<tr>
<td>Chang et al. (1997)</td>
<td>lime</td>
<td>cellulase: 5 FPU g(^{-1}) biomass, celllobiose: 28.4 CBU g(^{-1}) biomass</td>
<td>50 °C, 100 RPM, 72 h</td>
<td>85% conversion of biomass into reducing sugars</td>
</tr>
</tbody>
</table>

\(^a\) FPU, filter paper unit; IU, international enzyme activity unit; CBU, celllobiase unit

 instantaneous saccharification and fermentation (SSF), DMC involves the use of microorganisms that simultaneously produce cellulase to hydrolyze cellulose and ferment the resulting sugars into ethanol. Clostridium thermocellum and Clostridium thermosaccharolyticum have been used in DMC studies (Wyman, 1994). Significant by-product formation and low ethanol tolerance are limitations to this approach. In SSF, enzymatic hydrolysis and fermentation take place in the same vessel. The rationale for this approach is that since cellulase activity is inhibited by glucose, rapid fermentation into ethanol would increase the rate and efficiency of the overall process.

Wyman et al. (1992) reported significantly higher ethanol yields from dilute acid pretreated switchgrass using SSF over SHF. The study also noted ethanol yields from switchgrass were less than those from corn cobs and corn stover with SSF. Using a co-culture of Brettanomyces claussenii and S. cerevisiae, 87% (of theoretical maximum) conversion into ethanol using SSF was reported. Chung et al. (2005) reported comparable ethanol yields for SSF and SHF of dilute acid pretreated switchgrass (90.3% and 91.4%, respectively). Other studies have reported lower ethanol yields: 40% for SHF of AFEX pretreated switchgrass (Alizadeh et al., 2005) and 72% for SSF of lime-pretreated switchgrass (Chang et al., 2001). Reshamwala et al. (1995) used recombinant Klebsiella oxytoca to ferment both glucose and xylose in the enzymatic hydrolysate of AFEX pretreated switchgrass. Their results indicate that xylose fermentation was slower and less complete than glucose fermentation and greatly reduced at higher sugar concentrations. Fenske et al. (1998) used P. stipitis to ferment pentoses in the pretreatment liquor obtained from xylose-optimized dilute acid pretreatment of switchgrass and reported an 83% ethanol yield.

4. Thermal conversion

Combustion of biomass like switchgrass is problematic because of the presence of alkali metals that react to form sulfates, chlorides, silicate, and hydroxides that contribute to slag formation and fouling of combustion systems (Dayton et al., 1995). Leaching of switchgrass with water was proposed by Dayton et al. (1999) to remove alkali metals and chlorine prior to combustion. The study reported that leached biomass produced less alkali metal vapors and resulted in reduced char formation during combustion. Biomass cofiring with coal is a promising environmentally-friendly technology that has been shown to be economical with switchgrass replacing up to 20% of coal requirements (Tillman, 2000). Boylan et al. (2000) described plans for pilot scale testing of cofiring of switchgrass with coal. Preliminary experiments indicated a reduction in sulfur and nitrous oxide emissions for a mixture of 10% switchgrass–90% coal. Blevins and Cauley (2005) investigated the particulate matter formation during the combustion of coal and cofiring of coal-switchgrass mixtures and noted that fine particulate matter number densities were two orders of magnitude higher for mixtures. However, the particle mass concentrations remained unchanged. Brown et al. (2000) studied the synergistic effects during the co-gasification of coal-switchgrass and noted an almost 8-fold increase in the rate of coal char gasification.

Pyrolysis of Cave-in-Rock switchgrass to produce bio-oil was conducted by Boateng et al. (2007) using a fluidized bed system. The study reported bio-oil yields greater than 60% by mass with energy conversion efficiencies ranging from 52% to 81%. A previous study on pyrolysis of Cave-in-Rock switchgrass noted an interaction effect between pyrolysis temperature and plant maturity on profile of pyrolysis products (Boateng et al., 2006). Zhang et al. (2004) used switchgrass for thermochemical production of hydrogen. The process involved thermal gasification in a fluidized bed gasifier followed by steam reforming of tars in the producer gas. The high ratio of carbon monoxide to hydrogen was reduced by reacting carbon monoxide in the producer gas with steam. The study reported an increase in hydrogen concentration from 8.6% to 26.7% (by volume) in the producer gas.

5. Pulping and paper making

Although wood is the most common raw material for pulp- ing applications, environmental issues such as forest preservation and reduction of CO\(_2\) emissions have resulted in an interest in non-wood sources. Perennial grasses like switchgrass have low lignin content and can be harvested without annual reestablishment, making them an attractive raw material for pulp. Fox et al. (1999) conducted a regional economic evaluation for eastern Ontario and Western Quebec and concluded that switchgrass was an attractive crop for farmers and the local pulp industry. Vermeer et al. (2004) assessed pulping suitability of several plant materials and concluded that switchgrass could be used for producing paper for writing and printing or mixed with conventional woody sources for making paper with a wider range of applications. Madakadze et al. (1999) assessed pulping characteristics of different grasses and reported the highest pulp yield (51%) and pulp brightness (36%) for switchgrass. Although the study noted that fiber uniformity was an issue, the resulting paper had good printability. Law et al. (2001) reported pulp yields of 50–55% from soda-sulfite pulping of switchgrass and concluded that while brightness was low, the resulting pulp had excellent mechanical properties and could be used as a reinforcement component in newsprint production. In a later study, Law et al. (2002) reported pulp yields of 60–80%, but observed good mechanical strength only for switchgrass pulps in the yield range of 55–60%. Rozinski and Kokta (2000) also reported acceptable pulp quality using steam explosion pulping of switchgrass. Kraft pulp from switchgrass has short fibers with a high proportion of fines (Madakadze et al., 1999). Reddy and Yang (2007) used alkaline pretreatment to extract cellulose fibers from leaves and stems of switchgrass and noted tensile properties of fibers were similar to those of linen and cotton. Van den Oever et al. (2003) recognized that these fibers could be used as a reinforcing and filling agent for thermoplastic composites because of good fi-
ber quality and sustainable production at low cost. They noted that the addition of 30% (by weight) switchgrass pulp to polypropylene resulted in an increase of the flexural modulus by a factor of approximately 2.5 compared to pure polypropylene. This increase was only slightly lower than for jute and flax which are commonly used reinforcement fibers (Note: flexural modulus is an indicator of the ability of a material to resist deformation under load). The study concluded that optimization of the pulping process could further improve properties of the composite.

6. Conclusions and future prospects

Switchgrass is a promising feedstock for bioethanol production, thermal energy conversion and pulping applications. The positive environmental benefits associated with switchgrass include enhancement of wildlife diversity, improvement of soil and water quality, reduced pesticide use, and carbon sequestration. Most current research has focused on bioconversion into ethanol with an emphasis on pretreatment methods. While the technology for corn-to-ethanol is well established, the conversion of lignocellulosic feedstock to ethanol has challenges—such as cost of pretreatment methods, cost of hydrolytic enzymes, and inefficient fermentation of pentoses. The economics of bioethanol production from switchgrass can also be improved by developing value-added by-products. For example, Bals et al. (2007) present a scheme for extraction of proteins while simultaneously producing fermentable sugars from AFEX pretreated switchgrass. There is a lack of research on utilization of hemicellulose, which accounts for about 20–25% of switchgrass. With the development of genetically engineered microorganisms capable of efficient pentose fermentation, ethanol production from hemicelluloses in switchgrass needs to be optimized. Sarath et al. (2008) highlight opportunities in the area of genetics and agronomy to improve the biofuel characteristics of switchgrass. Along with increasing biomass yields, research on plant breeding should focus on modifying the composition of switchgrass to minimize recalcitrance to bioconversion. For example, a feedstock with lower levels of lignin can potentially reduce the severity and cost of the pretreatment required. However, the potential negative consequences on crop yields and fitness—as noted by Casler et al. (2002) and Pedersen et al. (2005) due to such modifications in biomass composition—need to be assessed against any potential benefits. While improvements in genetics, agronomy, and the conversion process will undoubtedly help in the development of a feasible biofuel production system from switchgrass, environmental and social stresses associated with dedicating large geographic areas for feedstock production needs to be addressed.

References


NREL Biomass Feedstock Composition and Properties Database, 2004; http://www.nrel.gov/biomass/energy_analysis.html


