The Importance of Demographic Niches to Tree Diversity

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Most ecological hypotheses about species coexistence hinge on species differences, but quantifying trait differences across species in diverse communities is often unfeasible. We examined the variation of demographic traits using a global tropical forest data set covering 4,500 species in 10 large-scale tree inventories. With a hierarchical Bayesian approach, we quantified the distribution of mortality and growth rates of all tree species at each site. This allowed us to test the prediction that demographic differences facilitate species richness, as suggested by the theory that a tradeoff between high growth and high survival allows species to coexist. Contrary to the prediction, the most diverse forests had the least demographic variation. Although demographic differences may foster coexistence, they do not explain any of the 16-fold variation in tree species richness observed across the tropics.

Comparative studies of tree demography typically consider the entire community as a unit, ignoring species differences (1), simply because most tree inventories include small samples of many species (2, 3). Comparative studies show that tropical forests typically have higher turnover than do temperate forests (4) and that higher tree turnover associates with higher tree diversity (5). These studies cannot, however, test ecological hypotheses about diversity, coexistence, and demography (6–10).

A tradeoff between rapid growth and long life span permits species coexistence and can foster diversity: Species reproducing early in life persist despite poor competition ability by growing rapidly on disturbed sites where resources are abundant. Long-lived species coexist by outliving the weedy invaders, persisting where resources are scarce. This is a familiar and widely known tradeoff in plant and animal communities (9–11) called the successional-niche hypothesis (7, 12). At a deterministic equilibrium, an indefinite number of species can coexist by this mechanism, each differing from all others along a continuum from short life span (with high growth) to long life span (and low growth). With stochastic demography, however, there is limiting similarity and the equilibrium species richness is finite (11, 13). This hypothesis is widely quoted as an explanation for tropical forest diversity (14–16). Here, we ask whether species differences along a demographic axis explain why some tropical forests have many more species than others.

If demographic niches are a key force controlling forest diversity, then more diverse forests have more demographic niches. More niches could come about either by spreading demographic rates over a wider range or packing more in the same range. Here, we focus on the first prediction: Tropical forests gain diversity by having a wider range of demographic niches, as reflected by the range of mortality and growth rates across species.

We provide a direct test by quantifying mortality and growth of 4,500 tree species in 10 different forests in America, Asia, and Africa (17). The 10 sites form a large-scale observation program, spanning a wide range of environmental conditions, designed to provide species-specific information for little-known tropical trees (18). At each site, a 20- to 52-ha tree census was set out in extensive, largely undisturbed forest (table S1). Species richness within the census plots differed by 16-fold, from 73 species per 50 ha in a dry forest at Mudumalai, India, to 1,167 species per 50 ha in a wet dipterocarp forest in Sarawak, Malaysia (19).

Past studies on the demography of individual tree species were based on direct measures of rate constants. These excluded many rare species because their rate estimates are subject to high error (20, 21). To overcome this limitation, we did not simply record species’ rates of mortality and growth; instead, we quantified the distribution of demographic rates across the entire community. A hierarchical Bayesian approach accomplishes this with explicit probability models covering both the observations of individual trees within species and the variation among species; all species, including rare ones, are included. For mortality, within-species distributions were modeled with the binomial distribution; for growth, we chose the log-normal based on the tendency for individual growth rates within a species to be highly skewed to the right. By separating within-species variation, the hierarchical model allows focus on the question of how species differ (10, 22, 23).

At the community level, we had to describe the variation in species’ demographic rates across species, and again, we used the log-normal to account for the skew to the right. Histograms of mortality rate μ and growth rate g (24) are fitted well by the log-normal when rare species are excluded (Figure 1). The log-normal requires two parameters, μ and σ; the mean and standard deviation of the natural logarithm of μ or g, respectively. We were able to estimate values of μ and σ that best describe a community’s demography with the use of the Gibbs sampler, simultaneously producing for every species estimates of mortality and growth rates that are adjusted for abundance. That is, for abundant species, the estimate is barely different from the observed rate, but for rare species, it is guided by the community-wide pattern (25).

Fitted log-normal distributions for the Lambir forest in Malaysia are plotted with observed histograms of sapling mortality and growth (Figure 1). The fit is close for more common species (filled bars), demonstrating that the large number of zeroes in the mortality histogram are sampling artifacts.
facts in rare species. Growth rates are also spread by rare species, though not as conspicuously. Fitted distributions for all the forests can be compared graphically (Figure 2) or with estimated 2.5 and 97.5 percentiles (Table 1). In the supporting online material, tables of mortality and growth rates for all species are provided (database S1).

Most of the forests were dominated by species with sapling mortality rates near 1% year⁻¹ (Figure 2). Even the high-mortality forests, such as Barro Colorado Island (BCI) and Hua Khae Khaeng National Park (HKK), had modes close to 1% year⁻¹ and low rates around 0.4% year⁻¹. The main feature separating these high-rate forests from the low-rate sites (such as Pasoh in Malaysia) is the long tail reaching 20% year⁻¹ mortality; at Pasoh, nearly all species had mortality rates of <3% year⁻¹. Thus, forests fell broadly in two groups: BCI, HKK, and La Planada had upper sapling mortality rates above 20%, whereas Sinharaja, Lambir, Pasoh, and Yasuni had upper rates below 8%. The Congo sites had exceedingly low mortality stretching to a modest 10% at the upper end.

Distributions of growth rates were similar to distributions of mortality, but growth was about half as variable across species (Figure 2). Conspicuously, sites with less variation in mortality also had less variation in growth (Figure 3). These patterns held for larger trees, although mortality and growth rates were lower (Table S2 and Figure S1).

Examples from individual species help illustrate. At BCI, a fast-growing understory treelet, *Palicourea guianensis*, had a population of 376 saplings in 1982, and every single one had died by 2005 (nevertheless, the population grew to 851). Although *Palicourea*’s mortality rate is infinite by direct calculation, the Gibbs sampler produces an estimate of 33% year⁻¹. *Alloplectus schultzei*, a small, weedy treelet at La Planada, also suffered 33% year⁻¹ mortality, losing 284 of 335 individuals over 6 years. In contrast, of 1,161 species at Lambir, none had mortality of 30% year⁻¹, and only two had rates above 20%; at Pasoh between 1987 and 2000, the very highest Gibbescorrected mortality rate among 802 species was 14% year⁻¹, in *Melastoma malabathricum*.

At the other end of the distribution, *Canpinia eglotica*, a midsized tree of the BCI understory, lost only 10 of 1,102 individuals between 1990 and 1995 (0.23% year⁻¹), and Carapa guianensis at La Planada, a large and valuable timber tree, lost only 11 of 894 (0.28% year⁻¹). In three census intervals at Pasoh, the lowest mortality rate recorded was 0.34% year⁻¹, in *Cynometra malaccensis*.
The Mudumalai forest stood out. Saplings had greatly elevated mortality and growth, with rates stretching much higher than any other site. Between 1988 and 1992, every species at Mudumalai had sapling growth of >6% year\(^{-1}\), and only BCI and HKK had many rates this high. At Lenda and Sinhara-jə, no species grew by 6% year\(^{-1}\). For larger trees (≥100 mm diameter), however, Mudumalai was in line with other forests, having modes of mortality and growth near 1% year\(^{-1}\) (Table S2). Indeed, trees at Mudumalai had among highest rates as well as the highest: Anogeissus latifolia had 116 deaths out of 2,179 trees from 1988 to 2000, whereas *Kydia calycina* had 1,272 of 1,328 trees die over the same interval, many because of elephant herbivory (26). Their rates differ by 50-fold: 27% year\(^{-1}\) in *Kydia* compared with 0.47% year\(^{-1}\) in *Anogeissus*.

Three of the sites with long tails of elevated mortality and growth—BCI, HKK, and Mudumalai—have intense annual dry seas-sons (Table S1). Mudumalai and HKK also burn in some years (26) (other plots do not burn and none suffer large-scale wind damage). It was not surprising that annual drought elevated mortality. Many species at these sites, however, had exceedingly low rates of mortality and evidently did not suffer much from drought. Conversely, forests lacking the tail of high growth and mortality had no or modest annual dry seasons, including the three forests dominated by Dipterocarpaceae (Sinhara-jə, Pasoh, and Lam-bir). Seasonality, however, was not the only factor predicting high variation in demography; the ever-wet cloud forest at La Plana-da, Colombia, had a wide spread of growth and mortality, comparable to the seasonal-ly dry sites.

Mudumalai and HKK have relatively open canopies compared with all of the other sites, a typical feature of dry forests, and many saplings at Mudumalai are sprouts from large root systems. These are likely reasons for elevated sapling growth at the two sites. In contrast, both BCI and La Plana-nada have dense canopies and dark understories, so canopy openness does not obvi-ously account for the high-growth species at those two sites.

Contrary to the prediction that demo-graphic variability begets species richness, diverse forests had the least variation in de-mography (Figure 3 and Figure S1). If any-thing, the most diverse forests had the few-est demographic niches. At Lambir, high species richness coupled with a low diversi-ty of demographic rates meant that 135 tree species coexisting in close proximity had sapling mortality rates in a narrow window from 0.8 to 1.0% year\(^{-1}\).

We do not question that demographic variability plays some role in species coex-is-tence. In American forests, the familiar ge-nus *Cocconoa* is found exclusively in small forest clearings (or outside the forest), where it rapidly colonizes and rapidly dies. The upper end of sapling mortality and growth distributions in America is set by gap spe-cialists: *C. obtusifolia* at BCI (12% year\(^{-1}\) mor-tality, 14% year\(^{-1}\) growth), *C. sciadophylla* at Yasuni (5.0% year\(^{-1}\) mortality, 6.3% year\(^{-1}\) growth), and *C. monstachya* at La Planada (8.8% year\(^{-1}\) mortality, 8.2% year\(^{-1}\) growth).

Diverse Southeast Asian forests lacked spe-cies with such high rates (27). The most diverse tropical forests are the least diverse demographically. It remains plausible that demographic niches are packed more tightly in some forests than others, but this seems unlikely, because packing should depend only on population size and turnover, which do not vary much. Moreover, the successional-niche hypothe-sis is not favored by the strong peak in de-mographic rates near 1 to 2% year\(^{-1}\); if demographic niches were crucial, then rates ought to be spread evenly over the entire range (28). Instead, the similarity in demog-raphy of many species suggests trait conver-gence (29). We believe that broad diversity differences are due to the source pool of dif-ferent biogeographic regions, and that demo-graphic differences play a minor role in species coexistence.

### Table 1. Variation in sapling mortality and growth rates across species in tropical forests

<table>
<thead>
<tr>
<th>Site</th>
<th>Years</th>
<th>No. of species</th>
<th>Median</th>
<th>Lower</th>
<th>Upper</th>
<th>Median</th>
<th>Lower</th>
<th>Upper</th>
</tr>
</thead>
<tbody>
<tr>
<td>BCI</td>
<td>82-85</td>
<td>284</td>
<td>3.14 ± 0.46</td>
<td>0.38 ± 0.10</td>
<td>26.0 ± 6.9</td>
<td>2.84 ± 0.16</td>
<td>1.35 ± 0.14</td>
<td>6.0 ± 0.7</td>
</tr>
<tr>
<td>BCI</td>
<td>85-90</td>
<td>282</td>
<td>2.56 ± 0.37</td>
<td>0.31 ± 0.08</td>
<td>21.5 ± 5.5</td>
<td>2.41 ± 0.18</td>
<td>0.85 ± 0.12</td>
<td>6.8 ± 1.0</td>
</tr>
<tr>
<td>BCI</td>
<td>90-95</td>
<td>282</td>
<td>2.85 ± 0.43</td>
<td>0.32 ± 0.09</td>
<td>25.3 ± 6.7</td>
<td>2.15 ± 0.13</td>
<td>0.89 ± 0.09</td>
<td>5.2 ± 0.7</td>
</tr>
<tr>
<td>BCI</td>
<td>95-00</td>
<td>285</td>
<td>3.31 ± 0.42</td>
<td>0.48 ± 0.11</td>
<td>23.3 ± 5.7</td>
<td>1.97 ± 0.12</td>
<td>0.94 ± 0.09</td>
<td>4.8 ± 0.6</td>
</tr>
<tr>
<td>BCI</td>
<td>00-05</td>
<td>285</td>
<td>2.91 ± 0.41</td>
<td>0.40 ± 0.10</td>
<td>21.4 ± 5.5</td>
<td>2.10 ± 0.16</td>
<td>0.73 ± 0.10</td>
<td>6.1 ± 0.9</td>
</tr>
<tr>
<td>Yasuni</td>
<td>96-03</td>
<td>1,077</td>
<td>1.55 ± 0.10</td>
<td>0.31 ± 0.04</td>
<td>7.7 ± 0.9</td>
<td>1.67 ± 0.04</td>
<td>0.83 ± 0.04</td>
<td>3.4 ± 0.2</td>
</tr>
<tr>
<td>La Plana</td>
<td>97-03</td>
<td>216</td>
<td>3.22 ± 0.47</td>
<td>0.45 ± 0.13</td>
<td>22.9 ± 5.9</td>
<td>2.30 ± 0.17</td>
<td>0.93 ± 0.13</td>
<td>5.7 ± 0.8</td>
</tr>
<tr>
<td>Pasoh</td>
<td>87-90</td>
<td>802</td>
<td>1.04 ± 0.06</td>
<td>0.38 ± 0.04</td>
<td>3.0 ± 0.3</td>
<td>1.50 ± 0.03</td>
<td>1.02 ± 0.04</td>
<td>2.5 ± 0.1</td>
</tr>
<tr>
<td>Pasoh</td>
<td>90-95</td>
<td>801</td>
<td>1.35 ± 0.07</td>
<td>0.42 ± 0.05</td>
<td>4.3 ± 0.4</td>
<td>1.59 ± 0.03</td>
<td>1.02 ± 0.04</td>
<td>2.5 ± 0.1</td>
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<tr>
<td>Pasoh</td>
<td>95-00</td>
<td>804</td>
<td>1.69 ± 0.09</td>
<td>0.47 ± 0.06</td>
<td>6.0 ± 0.6</td>
<td>1.55 ± 0.03</td>
<td>1.07 ± 0.04</td>
<td>2.3 ± 0.1</td>
</tr>
<tr>
<td>Lam-bir</td>
<td>92-97</td>
<td>1,161</td>
<td>1.32 ± 0.07</td>
<td>0.34 ± 0.03</td>
<td>5.2 ± 0.4</td>
<td>1.57 ± 0.03</td>
<td>0.96 ± 0.03</td>
<td>2.6 ± 0.1</td>
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<tr>
<td>HKK</td>
<td>93-99</td>
<td>296</td>
<td>4.11 ± 0.57</td>
<td>0.70 ± 0.19</td>
<td>24.1 ± 6.6</td>
<td>4.83 ± 0.45</td>
<td>1.53 ± 0.26</td>
<td>15.2 ± 2.7</td>
</tr>
<tr>
<td>Mudumalai 88-92</td>
<td>56</td>
<td>13.0 ± 3.48</td>
<td>6.25 ± 1.42</td>
<td>64.3 ± 36</td>
<td>7.87 ± 1.26</td>
<td>4.65 ± 1.52</td>
<td>13.3 ± 3.8</td>
<td></td>
</tr>
<tr>
<td>Mudumalai 92-96</td>
<td>52</td>
<td>17.06 ± 6.43</td>
<td>2.35 ± 1.51</td>
<td>124 ± 113</td>
<td>6.35 ± 1.67</td>
<td>2.57 ± 1.38</td>
<td>15.7 ± 11</td>
<td></td>
</tr>
<tr>
<td>Mudumalai</td>
<td>96-00</td>
<td>39</td>
<td>7.96 ± 2.70</td>
<td>1.73 ± 1.13</td>
<td>36.6 ± 25</td>
<td>5.71 ± 1.53</td>
<td>2.42 ± 1.69</td>
<td>13.4 ± 9.8</td>
</tr>
<tr>
<td>Sinhara-jə</td>
<td>95-01</td>
<td>205</td>
<td>1.35 ± 0.17</td>
<td>0.30 ± 0.07</td>
<td>6.0 ± 1.3</td>
<td>1.38 ± 0.07</td>
<td>0.75 ± 0.07</td>
<td>2.5 ± 0.2</td>
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<tr>
<td>Edoro</td>
<td>94-00</td>
<td>368</td>
<td>1.43 ± 0.20</td>
<td>0.21 ± 0.06</td>
<td>9.6 ± 2.5</td>
<td>1.41 ± 0.09</td>
<td>0.57 ± 0.07</td>
<td>3.5 ± 0.5</td>
</tr>
<tr>
<td>Lenula</td>
<td>94-00</td>
<td>346</td>
<td>1.26 ± 0.19</td>
<td>0.18 ± 0.06</td>
<td>8.8 ± 2.2</td>
<td>1.00 ± 0.04</td>
<td>0.64 ± 0.05</td>
<td>1.7 ± 0.1</td>
</tr>
</tbody>
</table>
References and Notes

2. Most forest censuses cover a single hectare or less and include only larger trees. Except in Europe and North America, such small samples capture less than half the local species, with many represented by a single individual.
17. Table 1 includes 4,956 species, but some occur at more than one site, and taxonomy has not been fully aligned across sites. We estimate 4,500 different species at the 10 sites.
19. The Sarawak plot covers 52 ha. For comparison, the number 1,167 was taken from a 50-ha subset.
24. The rate constant $m$ is the derivative of population size $N$ with respect to time, due to mortality alone. It is approximated by $\frac{\Delta N}{\Delta t}$. Relative growth rate is $g = \frac{\Delta \text{dbh}}{\Delta t}$, where dbh is the stem diameter at breast height.
25. Materials and methods are available as supporting material on Science Online.
29. S.P. Hubbell, Ecology, in press.
30. Errors in dbh measurements are 1 to 2 mm in saplings, and many saplings show slight negative growth (25). We, thus, cannot measure relative growth rates below about 0.5% year$^{-1}$. Where close to 0.5% year$^{-1}$, the low end of the sapling growth distributions are probably artifacts. In larger trees, errors are smaller relative to dbh, eliminating the problem.
31. Analyses were supported by NSF grant DEB-9806528 of the Research Coordination Network Program to the Center forTropical Forest Science. Data collection was funded by many organizations, principally the NSF, Andrew W. Mellon Foundation, Peninsula Community Foundation, Smithsonian Tropical Research Institute, Arnold Arboretum (Harvard), Indian Institute of Science, Forest Research Institute of Malaysia, Royal Thai Forest Department, National Institute of Environmental Studies (Japan), and John D. and Catherine T. MacArthur Foundation. We thank the hundreds of field workers who have measured and mapped trees.

Supporting Online Material: www.sciencemag.org/cgi/content/full/1124712/DC1

Materials and Methods
Figure S1
Tables S1 to S4
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Database S1
Computer Codes

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